



Traffic-related Air Pollution and the Right Ventricle

The Multi-ethnic Study of Atherosclerosis

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Abstract

Rationale: Right heart failure is a cause of morbidity and mortality in common and rare heart and lung diseases. Exposure to traffic-related air pollution is linked to left ventricular hypertrophy, heart failure, and death. Relationships between traffic-related air pollution and right ventricular (RV) structure and function have not been studied.

Objectives: To characterize the relationship between traffic-related air pollutants and RV structure and function.

Methods: We included men and women with magnetic resonance imaging assessment of RV structure and function and estimated residential outdoor nitrogen dioxide (NO₂) concentrations from the Multi-ethnic Study of Atherosclerosis, a study of individuals free of clinical cardiovascular disease at baseline. Multivariable linear regression estimated associations between NO₂ exposure (averaged over the year prior to magnetic resonance imaging) and measures of RV structure and function after adjusting for demographics,

anthropometrics, smoking status, diabetes mellitus, and hypertension. Adjustment for corresponding left ventricular parameters, traffic-related noise, markers of inflammation, and lung disease were considered in separate models. Secondary analyses considered oxides of nitrogen (NO_x) as the exposure.

Measurements and Main Results: The study sample included 3,896 participants. In fully adjusted models, higher NO₂ was associated with greater RV mass and larger RV end-diastolic volume with or without further adjustment for corresponding left ventricular parameters, traffic-related noise, inflammatory markers, or lung disease (all $P < 0.05$). There was no association between NO₂ and RV ejection fraction. Relationships between NO_x and RV morphology were similar.

Conclusions: Higher levels of NO₂ exposure were associated with greater RV mass and larger RV end-diastolic volume.

Keywords: air pollutants; pulmonary circulation; heart ventricles; pulmonary hypertension

Right heart failure is a cause of morbidity and mortality in obstructive and restrictive lung disease, left ventricular (LV)

dysfunction, and pulmonary arterial hypertension (1–3). Right ventricular (RV) hypertrophy is also associated with

increased risk for heart failure and cardiovascular death in community-dwelling adults without known cardiac

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At a Glance Commentary

Scientific Knowledge on the

Subject: Exposure to traffic-related air pollution has been linked to left ventricular hypertrophy, heart failure, and death. The lungs have substantial exposure to traffic-related pollutants; however, relationships between traffic-related air pollutants and right ventricular morphology have not been established.

What This Study Adds to the

Field: Higher levels of traffic-related air pollution, estimated by exposure to oxides of nitrogen, are associated with greater right ventricular mass and larger volumes. This relationship was not dependent on differences in left ventricular mass or volumes, systemic inflammation, roadway noise, or lung disease.

disease at baseline (4). Despite important epidemiologic and clinical roles of the right ventricle, little is known about modifiable determinants of RV structure and function (5).

Traffic-related air pollution is linked to LV hypertrophy, heart failure, and cardiovascular death (6, 7). Air pollution may affect the left ventricle through inflammation, oxidative stress, and autonomic dysfunction and these mechanisms could also affect the right ventricle (8–10). The lungs have substantial exposure to traffic-related air pollution and inhalants, which may directly increase RV afterload and lead to disproportionately greater changes in the right ventricle compared with the left ventricle (11, 12). The impact of traffic-related air pollution on the right ventricle, however, is not well-studied.

We examined the relationship between nitrogen dioxide (NO₂), a surrogate for traffic-related air pollution, and magnetic resonance imaging (MRI) measures of RV structure and function in a multiethnic cohort of adults free of clinical cardiovascular disease. We hypothesized that increased exposure to NO₂ would be independently associated with greater RV mass and larger RV end-diastolic volume (RVEDV). Some of the results in these studies have been previously reported in the form of an abstract (13).

Methods

The Multi-ethnic Study of Atherosclerosis (MESA) is a multicenter prospective cohort study designed to investigate subclinical cardiovascular disease in whites, African-Americans, Hispanics, and Chinese-Americans (14). Exclusion criteria included clinical cardiovascular disease (physician-diagnosed heart attack, stroke, transient ischemic attack, heart failure, angina, current atrial fibrillation, any cardiovascular procedure), weight greater than 136 kg (300 lb), pregnancy, or impediment to long-term participation. The Environmental Protection Agency funded a large ancillary study to MESA, the Multi-ethnic Study of Atherosclerosis and Air Pollution (MESA Air), which added cohort-specific air pollution monitoring and modeling (15). The MESA-RV study was an ancillary study funded to interpret cardiac MRIs for RV function. Individual participants gave informed consent and the institutional review boards of participating institutions approved the protocols of MESA and all studies described herein.

Traffic-related Air Pollution Exposure

Participants' residential address was assigned geographic coordinates using ArcGIS 9.1 software (ESRI, Redlands, CA) in conjunction with the Dynamap/2000 street network and geocoding database (Tele Atlas, Boston, MA). Using weighted averages of residential addresses over the year prior to cardiac MRI, individual outdoor home exposure to NO₂ and NO_x was estimated using spatiotemporal modeling and maximized by maximum likelihood (Figure 1) (16, 17). Estimates were fit using monitoring data from the Environmental Protection Agencies Air Quality System database and extensive cohort-specific air monitoring including home-based monitoring conducted as part of MESA Air (18). Geographic variables incorporated into the model included information on land use (e.g., industrial, residential); vegetative index; distance to various features (e.g., airports, coastline); road density; population density; elevation; urban topography; emissions sources; and dispersion model outputs integrating road position, traffic volume, diurnal traffic patterns, and meteorology.

Cardiac MRI Measures

Methods for acquisition and interpretation of LV and RV MRI parameters have been previously reported (19, 20). Endocardial and epicardial borders of the RV were manually traced on short axis cine images at end-systole and end-diastole. The outflow tract was included in RV volume. Papillary muscles and trabeculae were included in RV volumes and excluded from RV mass, as is commonly done for LV mass (21, 22). RV end-systolic volume and RVEDV were calculated using Simpson's rule by summation of areas on each slice multiplied by the sum of slice thickness and image gap. RV mass was determined at end-diastole as the difference between RV free wall end-diastolic epicardial and endocardial volumes multiplied by the specific gravity of the heart (1.05 g/ml). RV ejection fraction was calculated by subtracting RV end-systolic volume from RVEDV and dividing this difference by RVEDV.

Covariables

Covariables including age, sex, race/ethnicity, height, weight, education, income, presence of hypertension or diabetes mellitus, fasting plasma glucose, cholesterol, systolic blood pressure, smoking status and pack-years, percent emphysema (obtained by chest computed tomography), and self-reported lung disease (asthma and/or emphysema) were measured as previously described (23). Because levels of air pollution within a neighborhood are correlated over time, self-reported time a participant lived in the index neighborhood (the residential neighborhood used to determine 1-year pollutant estimates) was used as a surrogate for exposure duration (8). Participants reported roadway noise as a "very serious problem," "somewhat serious problem," "minor problem," or "not really a problem."

Statistical Analysis

We used linear regression to characterize relationships between NO₂ and RV parameters. All models were adjusted for height and weight, so it was not necessary to index RV parameters to account for differences in body size. Covariables were chosen *a priori* on the basis of known associations with ventricular size, heart disease, and comorbidities. In limited models, we adjusted for age, sex,

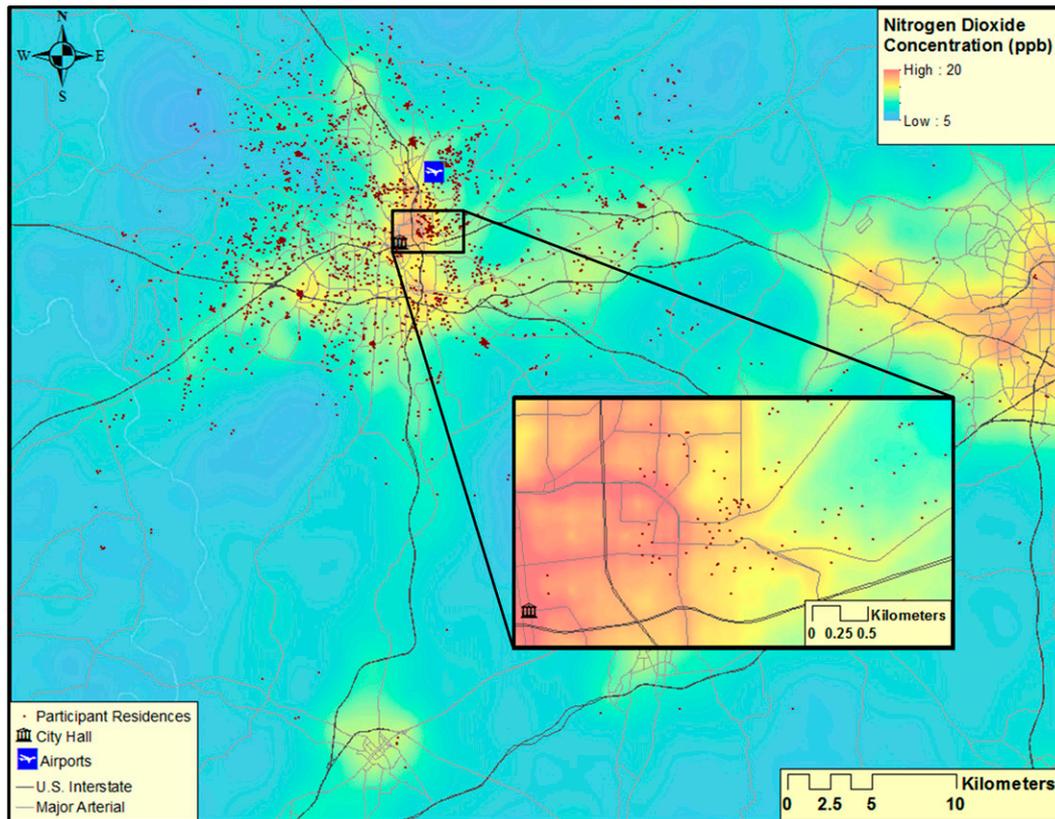


Figure 1. Representative map of Winston-Salem showing coarse and fine details of nitrogen dioxide predictions in parts per billion (ppb) from the spatiotemporal model including approximate Multi-ethnic Study of Atherosclerosis participant locations (jittered for privacy).

race/ethnicity, height, and weight (24). In fully adjusted models, we also included MESA field center; markers of socioeconomic status (self-reported income and education); and cardiovascular risk factors including smoking status, smoking pack-years, hypertension, cholesterol, diabetes mellitus, and impaired glucose tolerance. In prespecified models, we further adjusted for LV parameters, self-reported roadway noise, markers of inflammation (C-reactive protein and interleukin-6), or lung structure (% emphysema) and self-reported lung disease in separate models.

The primary analysis examined the relationship between RV parameters and NO_2 averaged over the year prior to cardiac MRI. Sensitivity analyses used fixed-year estimates of NO_2 in 2000, 2001, and 2002 to ensure there was no error introduced by the timing of the MRI in relation to secular exposure trends. Secondary analyses in limited and fully adjusted models used NO_x as the exposure of interest, which includes other components of the traffic-related air pollutant mix.

Several exploratory models further evaluated the relationship between NO_2 and RV metrics. Duration and timing of exposure were considered using a sliding time window analysis (25). We estimated associations between NO_2 and RV parameters in 5-year “time windows” (e.g., participants who lived in the index neighborhood for between 1 and 6 yr). The time window was then shifted by 1 year (e.g., participants who lived in the neighborhood between 2 and 7 yr) and new estimates of association and 95% confidence intervals (CI) were calculated. Overlapping 5-year periods avoid unstable estimates based on sparse data for a single calendar year and may more appropriately characterize the biologically relevant duration of exposure. Further exploratory models evaluated whether age, sex, or study site modified the association between NO_2 and RV parameters. We performed sensitivity analyses adjusting for body mass index category (normal weight and category 1–3 overweight) instead of height and weight to evaluate for residual confounding by obesity. Analyses were performed using STATA 12.0 (StataCorp, College Station, TX).

Results

There were 6,814 men and women enrolled in MESA (see Figure E1 in the online supplement) of whom 5,098 underwent cardiac MRI and 5,004 (98%) had interpretable examinations for the left ventricle. Of 4,634 participants selected for MESA-RV, MRI reads were attempted in 4,484 participants before achieving the study goal of 4,204 participants (94% of attempted reads). Outdoor exposure to NO_2 was estimated in 4,095 of these participants (97%). One hundred ninety-nine participants were excluded for missing covariables leaving 3,896 in the study sample. Table 1 shows characteristics of the study sample compared with those excluded. The mean age of the study sample was 61.4 years and 52.6% were women. Mean RV mass in the study sample was 21.1 ± 4.4 g, mean RVEDV was 124.2 ± 30.8 ml, and mean RV ejection fraction was $70.5 \pm 6.4\%$. Mean NO_2 was 21.8 ± 9.3 ppb with an interquartile range from 13.9 to 31.0 ppb. For individual cities the mean NO_2 ranged from

Table 1: Characteristics of the Study Sample Compared with Excluded Participants

	Study Sample (n = 3,896)	Excluded (n = 2,918)
Age, yr	61.4 ± 10.1	63.2 ± 10.4
Female, %	52.6	53.2
Race, %		
White	39.9	36.6
Chinese	12.5	10.8
African-American	25.6	30.6
Hispanic	22.0	22.0
Height, cm	166.4 ± 9.9	166.3 ± 10.2
Weight, kg	77.4 ± 16.2	80.3 ± 18.6
Body mass index, kg/m ²	27.8 ± 5.0	29.0 ± 6.0
Educational attainment, %		
No high school degree	15.8	21.1
High school degree	18.1	18.3
Some college	16.1	16.7
Bachelor's degree	18.5	15.5
Higher than bachelor's degree	19.1	16.5
Cigarette smoking status, %		
Never	52.6	47.3
Former	35.1	38.7
Current	12.4	14.0
Pack-years of smoking	10.8 ± 22.8	12.3 ± 21.4
Hypertension, %	42.5	48.4
Systolic blood pressure, mm Hg	125.3 ± 20.9	128.3 ± 22.1
Diabetes mellitus, %	12.3	15.3
Fasting plasma glucose, mg/dl	95.9 ± 28.2	99.3 ± 32.8
Study Site, %		
St. Paul	16.0	15.2
Los Angeles	18.2	20.9
Baltimore	17.7	13.6
Chicago	14.1	21.1
New York City	20.3	10.7
Winston-Salem	13.8	18.5
Stable residential neighborhood, %		
>5 yr	79.8	76.0
>10 yr	63.8	61.8
NO ₂ , ppb	21.8 ± 9.3	21.8 ± 8.6*
NO _x , ppb	50.5 ± 26.9	50.4 ± 26.7*

Definition of abbreviations: NO₂ = nitrogen dioxide; ppb = parts per billion. Data are shown as mean ± SD or percent when appropriate.

*A total of 1,055 participants with NO₂ and NO_x estimates not included in the study sample because of missing magnetic resonance imaging or covariables.

10.1 to 32.7 ppb and the city-specific interquartile range ranged from 3.1 to 5.0 ppb (see Figure E2).

Higher NO₂ was associated with greater RV mass (0.4 g for an interquartile increase in NO₂) (Table 2). This relationship became stronger after adjustment for city (0.9 g for an interquartile increase in NO₂) and after full adjustment for cardiovascular risk factors (1.0 g for an interquartile increase in NO₂) (Figure 2). This amounted to an approximately 5% increase in RV mass for an interquartile increase in NO₂. This significant association did not change with further adjustment for LV mass, traffic-related noise, inflammatory markers, or lung disease (Table 2; see Table E1).

Higher NO₂ was associated with larger RVEDV (2.9 ml for an interquartile increase in NO₂) (Table 2). This relationship became stronger after full adjustment for potential confounding by cardiovascular risk factors (4.1 ml for an interquartile increase in NO₂) (Figure 2). This amounted to an approximately 3% increase in RVEDV for an interquartile increase in NO₂. The significant association remained with further adjustment for LV end-diastolic volume, traffic-related noise, inflammatory markers, or lung disease (Table 2; see Table E1). NO₂ was not associated with RV ejection fraction (Table 2, Figure 2).

Secondary analyses using NO_x as the exposure of interest suggested relationships

similar to those for NO₂ but were in all cases modestly attenuated compared with NO₂ (see Table E3). RVEDV was not consistently associated with NO_x.

For participants with residential stability estimates (3,892 of 3,896 participants), sliding time window analyses indicated that participants who lived in the neighborhood several years before the MRI had incrementally stronger associations between NO₂ and RV mass than did those who lived in the neighborhood for a shorter duration (Figure 3; see Table E2). An incremental increase in RVEDV with participant duration in the neighborhood was less clear (Figure 3; see Table E2). Choice of the NO₂ reference period (calendar year 2000, 2001, or 2002) did not meaningfully impact the relationship between NO₂ and RV parameters (see Table E4).

Participant age did not modify relationships between NO₂ and RV parameters. The relationships of NO₂ with RV mass may have been stronger in men (1.3 g [95% CI, 0.4 to 2.2 g] per interquartile increase in NO₂) than women (0.6 g [95% CI, -0.1 to 1.3 g] per interquartile increase in NO₂) (*P* for interaction = 0.03). Similarly, the relationship of NO₂ with RVEDV may have been stronger in men (5.5 ml [95% CI, -0.3 to 11.2 ml] per interquartile increase in NO₂) than women (2.1 ml [95% CI, -2.2 to 6.5 ml] per interquartile increase in NO₂) (*P* for interaction = 0.04).

Participant city modified the relationship between NO₂ and RV mass (*P* for interaction < 0.001), but not RVEDV (*P* for interaction = 0.33). Qualitative associations between NO₂ and RV mass were in the same direction as the main association in St. Paul (6.4 g [95% CI, 4.1 to 8.8 g] per interquartile increase in NO₂), Los Angeles (0.9 g [95% CI, -0.1 to 1.9 g] per interquartile increase in NO₂), Baltimore (0.4 g [95% CI, -1.2 to 1.9 g] per interquartile increase in NO₂), and Chicago (0.3 g [95% CI, -1.0 to 1.6 g] per interquartile increase in NO₂). Qualitative associations were in the opposite direction as the main association in New York (-0.2 g [95% CI, -1.5 to 1.1 g] per interquartile increase in NO₂) and Winston-Salem (-0.4 g [95% CI, -2.6 to 1.8 g] per interquartile increase in NO₂). Because of the strong associations for St. Paul, we then excluded participants in cities with the greatest (St. Paul) and smallest (Winston-Salem) estimates of association between NO₂ and RV mass. The estimate of association in this four-city sample was smaller but qualitatively similar to the main analysis (0.5 g [95% CI, -0.1 to 1.1 g]

Table 2: Multivariable Linear Regression Estimating the Associations between NO₂ Exposure and Right Ventricular Structure and Function

Model	Per Interquartile Increase in NO ₂		
	Difference	95% CI	P Value
RV mass, g			
Limited model*	0.4	0.2 to 0.7	<0.001
Limited model* + city	0.9	0.3 to 1.4	0.002
Full model†	1.0	0.4 to 1.5	0.001
Full model† + LV mass	0.9	0.3 to 1.4	0.001
RVEDV, ml			
Limited model*	2.9	1.4 to 4.7	<0.001
Limited model* + city	2.7	−0.9 to 6.2	0.14
Full model†	4.1	0.5 to 7.7	0.03
Full model† + LVEDV	2.7	0.0 to 5.4	0.05
RVEF, %			
Limited model*	−0.1	−0.5 to 0.5	0.80
Limited model* + city	−0.2	−1.2 to 0.8	0.69
Full model†	−0.2	−1.2 to 0.8	0.72
Full model† + LVEF	0.0	−1.0 to 0.9	0.92

Definition of abbreviations: CI = confidence interval; LV = left ventricular; LVEDV = left ventricular end-diastolic volume; LVEF = left ventricular ejection fraction; NO₂ = nitrogen dioxide; RV = right ventricular; RVEDV = right ventricular end-diastolic volume; RVEF = right ventricular ejection fraction.

*Adjusted for age, sex, race/ethnicity, height, and weight.

†Adjusted for age, sex, race/ethnicity, height, weight, city, education, income, smoking status, pack-years, hypertension, diabetes, cholesterol, and impaired glucose tolerance.

increase per interquartile increase in NO₂; n = 2,738). Restricting this four-city sample to the sliding time window with the strongest association strengthened the relationship (1.3 g [95% CI, −0.1 to 2.7 g] increase per interquartile increase in NO₂; n = 476).

A sensitivity analysis adjusting for body mass index category, instead of the standard adjustment by height and weight, did not change the results of any analysis.

Discussion

We have shown that higher estimates of long-term outdoor residential NO₂ exposure are associated with greater RV mass and larger

RVEDV in a multiethnic, multicity cohort of adults without clinical cardiovascular disease. MESA participants had a 1.0 g (5%) increase in RV mass and 4.1 ml (3%) increase in RVEDV with an interquartile increase in NO₂. This difference in RV mass is quantitatively similar to that seen in LV mass in MESA participants with diabetes (2.4%) and in current smokers (5.3%), supporting biologic relevance (26, 27). RV hypertrophy in MESA participants is also associated with a three-fold increased risk of heart failure or cardiovascular death (4). This is the first report to suggest traffic-related air pollutants, of which NO₂ is a well-recognized surrogate for the pollutant mix, is associated with

morphologic changes in the right ventricle of the heart.

Our study provides initial insight into timing of this association. Duration of exposure to traffic-related air pollutants seems to be important. Participants who lived in the same neighborhood for several years had the strongest associations between NO₂ and RV mass. This suggests a dose–response, may provide insight for duration of necessary exposure, and supports a causal relationship.

The finding of both increased RV mass and RVEDV may suggest that the exposure of interest increased RV afterload (28). Previous studies have suggested that air pollution increases endothelin-1, a potent pulmonary vasoconstrictor (29), which could lead to increased pulmonary vascular resistance, increased RV afterload, and ultimately RV hypertrophy and dilation. Alternatively, air pollutants can irritate the respiratory epithelium and lead to heterogeneous ventilation with decreased regional ventilation (30). Regional hypoxia can cause hypoxic pulmonary vasoconstriction, increased resistance, and RV enlargement (31). Increases in afterload may compound oxidative stress and autonomic dysfunction, which have been implicated in the relationship between air pollution and LV mass and could directly contribute to RV pathology (8–10, 32).

Other mechanisms are possible as well. Air pollution may up-regulate myocardial inflammatory genes and proteins in the RV (33). Although it is not feasible to study myocardial gene and protein profiles in such a large study of the general population, our findings remained after adjustment for C-reactive protein and interleukin-6 blood levels, which suggests that our findings were independent of systemic inflammation. Roadway noise, which accompanies

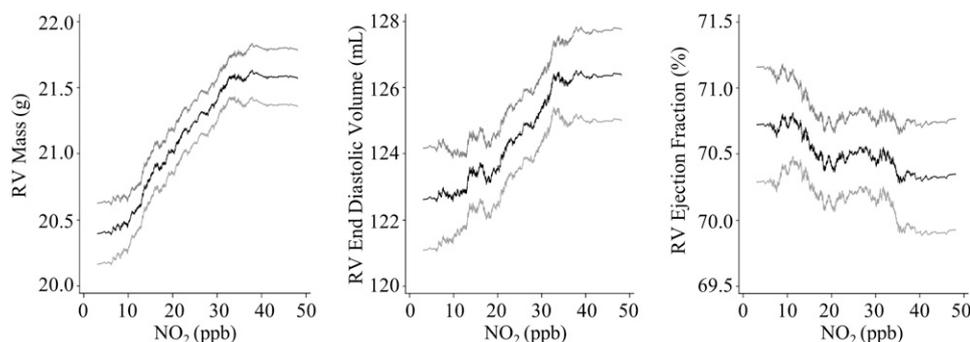


Figure 2. Multivariable nonparametric smoothed relationship between nitrogen dioxide (NO₂) in parts per billion (ppb) and right ventricular (RV) parameters with adjustment for age, sex, race/ethnicity, height, weight, city, education, income, smoking status, pack-years, hypertension, diabetes, cholesterol, and impaired glucose tolerance (black lines). Gray lines represent 95% confidence bounds.

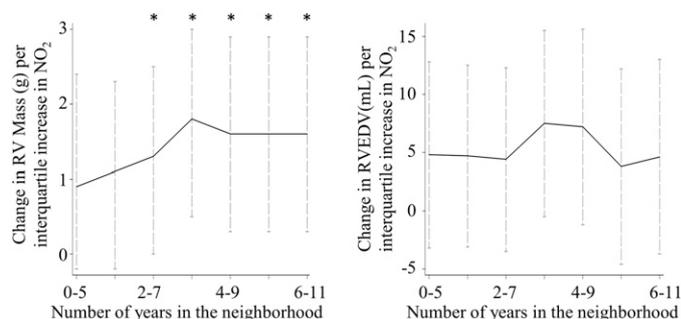


Figure 3. Relationship between the number of years a participant lived in their neighborhood and the difference in right ventricular (RV) mass or end-diastolic volume (EDV) per interquartile increase in nitrogen dioxide (NO_2): a sliding time window analysis of the full model. * $P \leq 0.05$.

traffic-related air pollution and may disrupt sleep, could mediate some aspects of the relationship between roadway proximity and heart disease (34). Adjusting for traffic-related noise did not attenuate relationships between NO_2 and RV morphology in our analyses.

Air pollution has also been linked to obstructive lung disease severity, which could increase RV afterload leading to increased RV mass (35, 36). However, we have previously shown that increasing airflow obstruction is associated with decreased RVEDV in MESA (37). In addition, adjustment for structural or self-reported lung disease did not change relationships between NO_2 and RV morphology in this analysis. Finally, LV mass may increase with traffic-related air pollution and LV hypertrophy can contribute to diastolic dysfunction and increased RV afterload, potentially explaining our results (6, 38). However, adjusting for the LV did not affect the results.

Relationships between NO_x and RV morphology were similar, but mildly attenuated compared with those with NO_2 . In addition, the relationship between NO_x and RVEDV was sensitive to adjustment. The NO_x analyses reinforce that the observed relationships are consistent with associations of a pollutant mix, not a specific pollutant, and that relationships between these pollutants and RV mass are stronger than relationships with RVEDV.

The association between NO_2 and RV mass was modified by city of residence. For

example, participants in New York City did not seem to have a relationship between NO_2 and RV mass despite the highest exposure to NO_2 . Heterogeneity by city is very common in air pollution research and was also seen in studies of LV mass and endothelial dysfunction (6, 39). Two key factors may contribute to city-specific heterogeneity. First, the validity of outdoor assessment of NO_2 as a surrogate for individual exposure to traffic-related pollutants depends on the degree to which outdoor pollution contributes to indoor pollution (e.g., home infiltration coefficient, indoor sources) and the proportion of time a participant spends indoors, outdoors, and in different microenvironments (40). These complex relationships vary among cities as a function of culture, climate, cooking/ventilation patterns, and average building age, among other factors. Second, our estimates of NO_2 are best conceptualized as a pattern of spatial decay consistent with some but not all traffic-related pollutants. For example, participants' exposure to NO_2 also reflects exposure to other hazardous air pollutants, such as benzene and several volatile organic compounds, levels of which may vary by city (41).

This study has limitations. Although we consider our exposure models to be a significant improvement over roadway proximity and nearest monitor analyses, measurement error and misclassification is likely present. Because error in exposure

assignments is unlikely to be dependent on RV measurements, these errors may be nondifferential with bias toward the null, so actual relationships may even be stronger than we have shown. Residual or unmeasured confounding, particularly at the neighborhood level, could contribute to the results. This may be especially true in the adjustment for exposure duration because neighborhood residents with long-term stability may differ from short-term residents. In the adjustment for road noise, measured or modeled noise would have been preferable to self-reports, but was not available. Furthermore, our study was cross-sectional and causality cannot be confirmed. Finally, measurement of invasive pulmonary hemodynamics, which may have informed the mechanism underlying our results, was not feasible in almost 4,000 community-dwelling participants free of cardiovascular disease.

Conclusions

Higher estimated exposure to NO_2 is associated with greater RV mass and larger RVEDV. This relationship is independent of markers of socioeconomic status, cardiovascular risk factors, left-sided cardiovascular disease, markers of inflammation, and lung disease. This is the first report to implicate traffic-related air pollution with changes in RV morphology. Air pollution may play a role in determining the RV response and outcomes in cardiopulmonary disease. ■

Author disclosures are available with the text of this article at www.atsjournals.org.

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